

Indirect land-use changes can overcome carbon savings from biofuels in Brazil

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The planned expansion of biofuel plantations in Brazil could potentially cause both direct and indirect land-use changes (e.g., biofuel plantations replace rangelands, which replace forests). In this study, we use a spatially explicit model to project land-use changes caused by that expansion in 2020, assuming that ethanol (biodiesel) production increases by 35 (4) $\times 10^9$ liter in the 2003–2020 period. Our simulations show that direct land-use changes will have a small impact on carbon emissions because most biofuel plantations would replace rangeland areas. However, indirect land-use changes, especially those pushing the rangeland frontier into the Amazonian forests, could offset the carbon savings from biofuels. Sugarcane ethanol and soybean biodiesel each contribute to nearly half of the projected indirect deforestation of 121,970 km² by 2020, creating a carbon debt that would take about 250 years to be repaid using these biofuels instead of fossil fuels. We also tested different crops that could serve as feedstock to fulfill Brazil's biodiesel demand and found that oil palm would cause the least land-use changes and associated carbon debt. The modeled livestock density increases by 0.09 head per hectare. But a higher increase of 0.13 head per hectare in the average livestock density throughout the country could avoid the indirect land-use changes caused by biofuels (even with soybean as the biodiesel feedstock), while still fulfilling all food and bioenergy demands. We suggest that a closer collaboration or strengthened institutional link between the biofuel and cattle-ranching sectors in the coming years is crucial for effective carbon savings from biofuels in Brazil.

deforestation | integrated assessment | livestock | policy analysis | Global Environment Outlook 4

Brazil's government and biofuel industry are planning a large increase in the production of biofuels in the next 10 years. This increase is driven by internal and external market demand (ethanol), as well as by government-enforced blending (biodiesel) (1–3). Although Brazilian sugarcane ethanol is often considered to have one of the best production systems with respect to carbon savings (4–8), there are concerns about the land-use changes (LUC) that would be incurred by an expansion of biofuel croplands (6, 7). Soybean plantations, from which most of the Brazilian biodiesel is produced (1, 3), already occupy 35% of the country's cultivated land (9). It is known that biofuels can replace vast areas of farmland and native habitats, driving up food prices and resulting in little reduction of or even increasing greenhouse gas (GHG) emissions (6, 7, 10–15).

Previous studies focused on the direct land-use changes (DLUC) and the “carbon debt” caused by the replacement of native habitats by biofuel crops in Brazil (7, 8, 10, 11). Others pointed to the probable indirect land-use changes (ILUC) in Brazil caused by future expansion of biofuel croplands in the United States (14–16). Overall, these studies show that potential LUC must be taken into account to assess the efficacy of a given biofuel. However, these studies were neither spatially explicit, nor did they explicitly consider competition between different land uses in view of concurrent food and biofuel demands. Fargione et al. (7), for example, show the LUC carbon debt in terms of rate (e.g., MgCO₂ ha⁻¹), since they did not consider the total extent of land dedicated to biofuels or the

total area of native habitats affected. Therefore, the net debt in absolute terms (e.g., MgCO₂) arising from future biofuel production remains undetermined. Moreover, the cascade effect of biofuel crops pushing the agricultural and cattle ranching frontier is still poorly understood.

Most of Brazil's sugarcane expansion in the last 5 years occurred on land previously used as rangeland in the southeastern states (11, 17). The same holds true for more than 90% of the soybean plantations in the Amazon region after the 2006 moratorium was implemented (18). One of the potential consequences of such LUC is the migration of cattle ranchers to other regions and possible increased deforestation (16, 19–21). In light of the role rangeland plays in deforestation in Brazil (16, 19–21) and the steadily increasing cattle herd [average of 3 million additional head per year in the 1974–2007 period (9)], the ILUC to replace rangeland displaced by biofuels are highly important (22).

In this study we use a spatially explicit modeling framework to project the DLUC and ILUC arising from the fulfillment of Brazil's biofuel production targets for 2020 concurrent with increasing food and livestock demands. This modeling framework comprises: (i) a land-use/land-cover change model for land-use suitability assessment and allocation (23); (ii) a partial equilibrium model of the economy of the agricultural sector for future food and livestock demands as well as technological improvements of crop yields (24); and (iii) a dynamic global vegetation model for crop and grassland potential productivity driven by climate (25, 26). Competition among land uses (for land resources) is considered based on a multicriteria evaluation of suitability, hierarchical dominance of major land-use activities (settlement, crop cultivation, grazing), and a multiobjective land allocation algorithm which looks for land-use pattern stability. Final outputs of this modeling framework are maps of land use and livestock density (Ld). DLUC and ILUC are determined by comparing land-use maps derived from scenarios with and without biofuel expansion. A number of different scenarios are considered to assess the isolated contribution of ethanol and biodiesel fuel production, as well as their impacts on different native habitats. The carbon debt and payback time from such LUC are calculated by using the average emission values employed by Fargione et al. (7). We investigate only the effects of ILUC inside Brazil. We do not consider cellulosic biofuels because the technological development of these fuels is unlikely to be fast enough to enable their large scale use in Brazil by 2020 (27).

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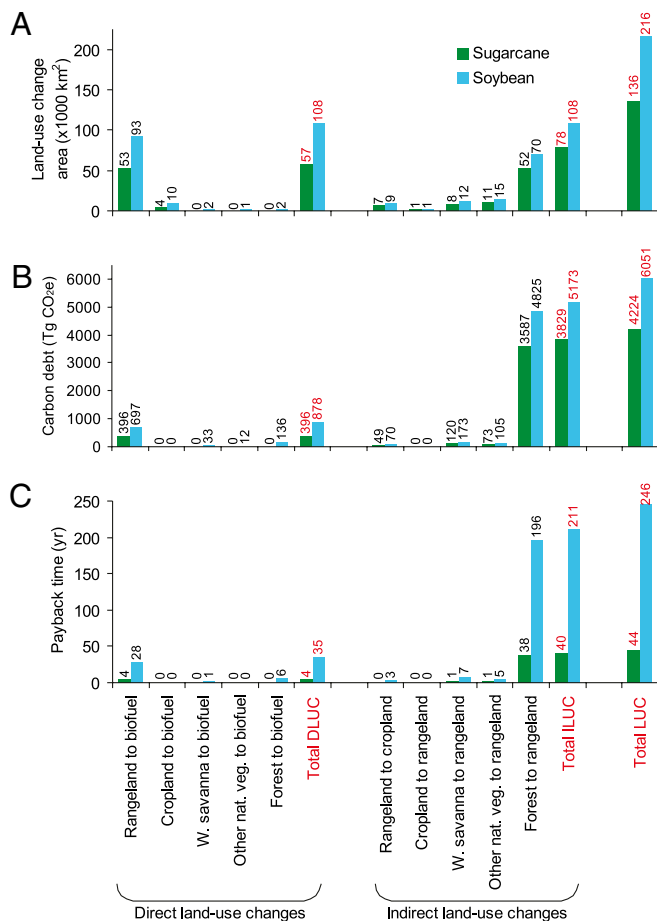


Fig. 2. Direct, indirect, and total LUC areas (A), carbon debt (B), and time to repay debt (C) for fulfilling Brazil's biofuel (sugarcane ethanol and soybean biodiesel) production targets for 2020. Here the land-use category "cropland" excludes sugarcane and soybean. Other nat. veg., other natural vegetation; W. savanna, woody savanna.

achieve such a reduction in rangeland area but still meet the same livestock demand, Ld would need to be increased by 0.13 head per hectare in comparison with 2003 values.

Other Biodiesel Feedstocks. It can be argued that soybean is not the most efficient feedstock for biodiesel because it occupies large tracts of land, incurs considerable carbon debt (even without considering ILUC), and has a low annual rate of saved carbon from replacing fossil diesel. Therefore, we tested other feedstock options that could serve to fulfill Brazil's 2020 production demand for biodiesel. Our

results show (Fig. 3 and Fig. S2) that if the smallest area and carbon debt from LUC are given priority, then oil palm would be the best feedstock for biodiesel by far. Because of its high oil yield, oil palm would need only 4,200 km² to fulfill the 2020 demand for biodiesel in Brazil. In comparison, 108,100 km² would be needed for soybean, 73,000 km² for rapeseed/sunflower, and 31,700 km² for *Jatropha curcas*. The payback time for oil palm would be 7 years for DLUC, which is much smaller than the DLUC payback time of 27 years for sunflower/rapeseed. However, if oil palm is strictly planted only in rangeland areas, the DLUC payback time would be reduced to 4 years. Sunflower/rapeseed plantations would be located mainly in south-central states. *Jatropha curcas* plantations, which were forced to occur only in Northeast Brazil in our simulations (Methods), are concentrated in the coastal area, where potential yields are higher. Oil palm plantations would be located entirely in Pará state, close to the Amazon forest and where most current plantations are located (9). Oil palm would incur some direct deforestation (300 km²), although much less than that directly caused by soybean. If oil palm is used as biodiesel feedstock in conjunction with sugarcane for ethanol, then Ld would need to be increased by only 0.10 head per hectare from 2003 to 2020 to avoid ILUC, compared to the 0.13 head per hectare increase needed for the soybean-sugarcane combination.

Discussion

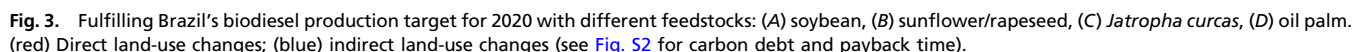
Our results show that sugarcane-ethanol and oil palm-biodiesel grown in Brazil are the best plant feedstocks in terms of carbon savings for fulfilling the country's demand for biofuels in 2020, assuming that the LUC associated with the increased production are restricted to the DLUC in rangelands. The simulated DLUC, which occur predominantly in rangelands, have already been observed for sugarcane (9, 11, 17) and soybean (16, 18, 19) in recent years. For sugarcane, this trend will probably continue in the next years because of the growing number of standards being imposed on sugarcane plantations (2, 28). However, the proximity of sugarcane plantations to Atlantic forest remnants in Southeast Brazil is of particular concern, considering that any further deforestation there would have major impacts on the biodiversity and connectivity of this highly threatened forest (29). The moratorium on soybean introduced in 2006 has proven to be an efficient way for preventing deforestation directly caused by soybean production in the Amazon region (18). Moreover, increasing pressure by the media and nongovernmental organizations (30) suggests that the moratorium will continue to be respected in the coming years. Even though oil palm is strongly associated with deforestation in Southeast Asia (31), the Brazilian palm oil production is still small and could be expanded into nonforest sites, assisted by improved governance in the Amazon region (32, 33).

The efficacy of biofuels in Brazil can be considerably compromised if biofuel-related ILUC, namely moving the rangeland frontier into native habitats, take place as projected here. It has been suggested that ILUC indeed occur in the Amazon region, especially the case where rangeland is shifted by soybean and

Table 1. Land-use and land-use change (relative to 2003) according to different modeled scenarios for Brazil in 2020

	Soybean	Sugarcane	Other crops	Rangeland	Forest	Other natural	Livestock density
Scenario	x1,000 km ² (Δ03–20, %)	x1,000 km ² (Δ03–20, %)	x1,000 km ² (Δ03–20, %)	x1,000 km ² (Δ03–20, %)	x1,000 km ² (Δ03–20, %)	x1,000 km ² (Δ03–20, %)	head/ha (Δ03–20, %)
2003	191 (—)	55 (—)	389 (—)	2,133 (—)	4,194 (—)	1,496 (—)	0.70 (—)
2020 biofuel targets	285 (+49.3)	90 (+63.6)	397 (+2.1)	2,972 (+44.3)	3,546 (–15.4)	1,222 (–21.6)	0.79 (+12.3)
2020 no biofuel expansion	178 (–6.9)	33 (–39.9)	398 (+2.4)	2,968 (+44.1)	3,668 (–12.5)	1,268 (–18.7)	0.79 (+12.5)
2020 ethanol targets only	178 (–6.8)	90 (+62.8)	398 (+2.4)	2,973 (+44.3)	3,618 (–13.7)	1,255 (–19.5)	0.79 (+12.3)
2020 biodiesel targets only	285 (+49.5)	33 (–39.7)	396 (+2.0)	2,968 (+44.1)	3,598 (–14.2)	1,232 (–21.0)	0.79 (+12.5)
2020 biofuel targets no ILUC	285 (+49.3)	90 (+63.6)	397 (+2.1)	2,807 (+36.2)	3,668 (–12.5)	1,268 (–18.7)	0.83 (+18.9)

ha, hectare (equivalent to 1 hm²).



Finally, the efficacy of biofuels is analyzed here in terms of GHG savings and not from the socioeconomical perspective. As a counterpart to the ethanol production chain, Brazil's National Program on Biodiesel Production seems to aim at promoting small-scale farming and shortening dependence on conventional diesel (3). However, between 75 and 95% (depending on the year) of the biodiesel produced in Brazil so far comes from soybean grown on plantations that are owned or controlled by large-scale farmers, and at production costs that are higher than for production of fossil diesel (1, 16, 19). Comprehensive assessments of labor conditions, land division, food prices, and other socioeconomical implications arising from the expansion of biofuels in Brazil are yet to be done. Nevertheless, joining life-cycle assessment figures to spatially explicit LUC projections, like the present study does, allows for a more accurate evaluation of the efficacy of biofuels in terms of carbon savings.

Methods

Modeling Framework. The central part of our modeling framework is the LandSHIFT model, which simulates land-use and land-cover change in a spatially explicit way at a resolution of 5 arc minutes (23). The model relies on a “land-use systems” approach that describes the interplay between anthropogenic and environmental system components as drivers for LUC in three major land-use activities (settlement, crop cultivation, and grazing) and their competition for land resources. It calculates not only the occurrence of grazing but also the intensity at which it occurs. LandSHIFT has been applied and validated in assessments of the impact of grazing management in the Jordan River region (40) and quantification of future LUC and water use by agriculture in Africa (41). The framework also comprises other models that, although not coupled to LandSHIFT, provide inputs to the model. The International Model for Policy Analysis of Agricultural Commodities and Trade (IMPACT) (24) calculates future country-level food demands and technological improvements of crop yields, and the International Futures model (42) projects population growth. The LPJ for managed Lands (LPJmL) dynamic global vegetation model is used to calculate crop and grassland potential productivity on a 0.5° resolution grid (25, 26). Starting from an initial land-use map, the spatial allocation of different land uses in subsequent time steps is based on a multicriteria-suitability analysis following the equation:

$$\mu_k = \underbrace{\sum_{i=1}^n w_i p_{i,k}}_{\text{suitability}} \times \underbrace{\prod_{j=1}^m c_{j,k}}_{\text{constraints}}, \text{ with } \sum_i w_i = 1, \text{ and } p_{i,k}, c_{j,k} \in [0, 1] \quad [1]$$

where the factor-weights w_i determine the importance of each suitability factor p_i at grid cell k , and c_j determines constraints for changing the land-use type at that given cell. In this study, p_i includes potential crop/grassland yield, slope, proximity to settlements, proximity to cropland, road network, and soil fertility (the latter does not apply for grazing). Therefore, $n = 6$ (5 for grazing). The weights w_i for cropland were determined with the use of the analytic hierarchy process test (43). In the *SI Text* we describe the determination of the relative importance of each p_i factor in relation to the others, which is used as an entry to the analytic hierarchy process test. Weights for road network, slope, and soil were fine-tuned from 0.23 to 0.13, from 0.18 to 0.23, and from 0.23 to 0.29, respectively (Table S3), to improve spatial distribution of croplands inside the country. The weights w_i for rangeland were assigned all of the same value (0.2). Constraints c_j are applied in cells that are designated as conservation areas or according to the land use transition in question (Table S4). A third “constraint” was implemented for sugarcane and soybean. This constraint represents the preferential occurrence of these crops in places where specific infrastructure is or will be implemented [such as ethanol mills (5)] or as in the case of soybean, where production costs are lower (3) and there is political facilitation for the cultivation of soybean (16, 19). The suitability for sugarcane is increased by 35% in the states of São Paulo, Minas Gerais, Mato Grosso do Sul, Goiás, and Distrito Federal. Suitability for soybean is increased by 35% in the states of Goiás, Tocantins, Mato Grosso do Sul, and Distrito Federal. In Mato Grosso, suitability is increased by 80%. These values were chosen to better reproduce the area of sugarcane and soybean in these states (Model Evaluation in the *SI Text*). The allocation algorithm assumes that crop cultivation takes place on the most suitable cells for each crop type and calculates a “quasi-optimum” spatial crop distribution. However, the multiobjective land allocation heuristic used here seeks pattern stability and respects previous land use, even if another crop type has a higher suitability in that cell (Table S4). Besides soybean and sugarcane, nine other major crop types are considered, including maize, pulses, rice, and wheat. LPJmL yields are applied a crop-specific factor to match current crop yields to statistics on the country level (Table S1) (23, 44). These factors, which are calculated at the first simulation time step, account for uncertainties because of crop management, (e.g., multicropping), or discrepancies because of the aggregation of crop types into the LPJmL crop functional types (e.g., LPJmL pulses refer to extra-tropical pulses, such as lentils). Crop production of a given grid cell k is defined as the potential crop yield at k multiplied by the area in k that is not covered by settlement.

Allocation of rangeland relies on the potential productivity of grass in the grid cells, based on a livestock feed supply-demand logic. Forage supply is calculated by summing up the grass productivity of every rangeland cell multiplied by the fraction of biomass that can be used by livestock [grazing efficiency $g_e = 0.3$ (45)]. Forage demand is determined by the multiplication of the total livestock herd by the average forage consumption per livestock unit [4.6 Mg/yr (46, 47)]. We assume that 95% of the livestock demand is fulfilled by forage from pastures (47). If forage demand is higher than the supply, then new rangeland cells are allocated, starting from grid cells with higher suit-

ability and continuing until demand is fulfilled. Average Ld is calculated by dividing the total livestock herd by the rangeland area. Preferential allocation of land-use activities follows the order: settlement, crop cultivation, grazing. Only one dominant land-use type can occur in a grid cell.

Input Data and Modeling Protocol. LandSHIFT is initialized with a combined map of land cover and land use for the year 1992 (48), a map of population density (49), and national statistics of crop production and livestock herd (44). Socioeconomic projections include future demands for food production, technological improvements of crop yields (24), and population growth (42) generated for the United Nations Environment Programme's Global Environmental Outlook 4 (GEO4) report under the Sustainability First scenario (50). We focus our analysis on this scenario because it predicts the highest use of biofuels worldwide by far and the largest increase in food production in Brazil (50). For the sake of scenario consistency, future potential crop and grassland productivity was calculated with the LPJmL model (25, 26) in 0.5° spatial resolution using as input a climatology of temperature, precipitation, and $[CO_2]$ from the IMAGE model (51), which was also generated for the GEO4 report (50). The Sustainability First scenario used here depicts a global mean increase in temperature of 1.1 °C in 2020 in relation to preindustrial times and an atmospheric CO_2 concentration of 426 ppm (50). There is a national population increase from 177 million people in 2003 to 202 million in 2020, with an average growth rate of 1.62% per year (42). Oil palm yields, which are not modeled by LPJmL, are simulated by applying a factor of 6.0 to the yields of the tropical roots crop functional type. Resulting yields are in accordance with oil palm yields in Brazil from census data for the 1990s (9) (Northeastern Pará: simulated = 11.4 Mg/ha, census = 13.6 Mg/ha; eastern Bahia: simulated = 3.3 Mg/ha, census = 4.1 Mg/ha; roughly 99% of Brazil's oil palm area is located in these two regions). Average potential yields for the 1990s are used as baseline yields in LandSHIFT (Table S1). On average, food production increases by 86% in the 2003 to 2020 period (Table S2), and yields increase, on average, by 62% because of the combined effects of technological improvements and climate change (Table S1). Livestock herd grows from 149 million head in 2003 to 234 million head in 2020, with an annual increase of 3.4%. This increase rate is slightly larger than the 3.25% average annual growth rate observed over the last 30 years (9).

Biofuel production follows the official projections by the Brazilian government and the biofuel industry (1, 2) (Table S5). Demands for food and biofuels are fed separately into the model, but they are treated equally inside the model algorithm. No preference is given to either food or biofuels. The main claims for using *Jatropha curcas* as a biodiesel feedstock are its drought tolerance, the low management inputs needed for its cultivation, and the inclusion of small farmers in the production chain (26), which is in accordance to Brazil's National Program on Biodiesel Production (3). For that reason, we restrict the occurrence of *J. curcas* to Northeast Brazil, which is the region targeted by the Brazilian government for inclusion of small-scale farming (3). This restriction is not applied to the other feedstocks.

Four scenario variations are modeled: (i) biofuel targets: 2020 food + 2020 biofuel production; (ii) no increase in biofuel production: 2020 food + 2003 biofuel production; (iii) ethanol targets only: 2020 food + 2020 ethanol + 2003 biodiesel production; and (iv) biodiesel targets only: 2020 food + 2003 ethanol + 2020 biodiesel production. DLUC are determined by the changes in the area covered by biofuel crops in variation (i) compared to variation (ii). ILUC are determined by the difference in the area covered by land uses other than biofuel crops between variations (i) and (ii). The intensification of livestock needed to avoid ILUC by biofuels is estimated by increasing the grazing efficiency (g_e) factor to the level at which rangeland area is equal to that of variation-scenario (i) minus the area of rangeland displaced by biofuels.

Model Evaluation. LandSHIFT model results for Brazil were evaluated in three aspects (a detailed presentation is given in the *SI Text*).

Crop and rangeland location. A comparison of LandSHIFT's calculated suitability with a reality land-use map showed a tendency for the occurrence of high suitability values in crop and rangeland, suggesting the procedure used in LandSHIFT is reasonable for allocation of crop and rangeland (Fig. S3). The land-use map used in this comparison is the same as the one used in the determination of the w_i weights, generating a spurious dependency between the datasets used for comparison. However, an analysis where all w_i weights were set to the same value of 0.16 further confirmed the tendency of high-suitability values in crop and rangeland grid cells. A second test using the relative operating characteristics (ROC) method (52) showed that the spatial pattern computed by LandSHIFT [ROC (cropland) = 0.87; ROC (rangeland) = 0.80] is not random, in which case it would have a value of 0.5 (Fig. S4).

Crop and rangeland area. Modeled cropland and rangeland areas are in very good agreement with country-level reported statistics (44), suggesting the model is able to convert country-scale crop production mass into cropland

area (Fig. S5). Overestimation of rangelands by 8% might be the result of an underestimation of grassland productivity and also because of the assumption of only one land use per grid cell, which leads to overestimation of the rangeland area, especially in regions where L_d is low, as in Northeast Brazil. Crop and rangeland areas within major regions of Brazil are also in good agreement with statistics (9), except for the overestimation of rangeland in Northeast Brazil (Fig. S6).

Deforestation rates. The modeled annual deforestation rate for the Amazon region in the 1992 to 2003 period compares well with remote sensing data [LandSHIFT: 16,789 km²/yr, INPE-PRODES: 18,266 km²/yr (53)]. The shares of this deforestation among states are also comparable with PRODES, even though deforestation in Maranhão is overestimated by a factor of 23. However, the land-use map used for model initialization [based on IGBP-DISCover dataset (48)] has 80% more forest in Maranhão compared to PRODES (53). Moreover, LandSHIFT does not consider forestry activities, which might influence deforestation rates. The modeled deforestation rate

of Central Brazil Cerrado for the 1992 to 2003 period is 17,753 km²/yr, an amount that lies within the estimated range for the last decade [13,100–26,000 km²/yr (12)].

Carbon Debt and Payback Time. Carbon debt and payback time are calculated following the approach used by Fargione et al. (7), with two major differences. First, the final numbers are absolute values (Mg CO₂) rather than rates (Mg CO₂ ha⁻¹) because we calculate total LUCs, and second, the annual CO₂ offset by biofuels are calculated on a per ton basis instead of a per hectare basis. All numbers used in the carbon debt and payback time calculations are shown in Tables S6 and S7.

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